Provided for non-commercial research and education use. Not for reproduction, distribution or commercial use.



(This is a sample cover image for this issue. The actual cover is not yet available at this time.)

This article appeared in a journal published by Elsevier. The attached copy is furnished to the author for internal non-commercial research and education use, including for instruction at the authors institution and sharing with colleagues.

Other uses, including reproduction and distribution, or selling or licensing copies, or posting to personal, institutional or third party websites are prohibited.

In most cases authors are permitted to post their version of the article (e.g. in Word or Tex form) to their personal website or institutional repository. Authors requiring further information regarding Elsevier's archiving and manuscript policies are encouraged to visit:

http://www.elsevier.com/copyright

Aquatic Toxicology 122-123 (2012) 36-43

Contents lists available at SciVerse ScienceDirect



Aquatic Toxicology



journal homepage: www.elsevier.com/locate/aquatox

The effect of nonylphenol on gene expression in Atlantic salmon smolts

Laura S. Robertson^{a,*}, Stephen D. McCormick^b

^a USGS, Leetown Science Center, 11649 Leetown Road, Kearneysville, WV 25430, USA ^b USGS, Leetown Science Center, Conte Anadromous Fish Research Center, Turners Falls, MA 01376, USA

ARTICLE INFO

ABSTRACT

Article history: Received 8 March 2012 Received in revised form 15 May 2012 Accepted 24 May 2012

Keywords: Endocrine-disrupting compound Expression Microarray Nonylphenol Salmon Smolt The parr–smolt transformation in Atlantic salmon (*Salmo salar*) is a complex developmental process that culminates in the ability to migrate to and live in seawater. Exposure to environmental contaminants like nonylphenol can disrupt smolt development and may be a contributing factor in salmon population declines. We used GRASP 16K cDNA microarrays to investigate the effects of nonylphenol on gene expression in Atlantic salmon smolts. Nonylphenol exposure reduced gill Na⁺/K⁺-ATPase activity and plasma cortisol and triiodothyronine levels. Transcriptional responses were examined in gill, liver, olfactory rosettes, hypothalamus, and pituitary. Expression of 124 features was significantly altered in the liver of fish exposed to nonylphenol; little to no transcriptional effects were observed in other tissues. mRNA abundance of genes involved in protein biosynthesis, folding, modification, transport and catabolism; nucleosome assembly, cell cycle, cell differentiation, microtubule-based movement, electron transport, and response to stress increased in nonylphenol-treated fish. This study expands our understanding of the effect of nonylphenol on smolting and provides potential targets for development of biomarkers. Published by Elsevier B.V.

1. Introduction

Historically, the North American range of Atlantic salmon (*Salmo salar*) extended southwards along the eastern United States to Long Island and the Housatonic River (Parrish et al., 1998). Now, all natural populations in the Long Island Sound and Central New England regions have been extirpated (Gephard and McMenemy, 2004). Exposure to environmental contaminants may be a contributing factor in these declines. Forest-spraying of pesticides that use nonylphenol as a surfactant are correlated with historical declines in adult return rates of Atlantic salmon in Canada (Fairchild et al., 1999), and its population level effects were hypothesized to be through lower survival of smolts after seawater entry. Nonylphenol has also been shown to inhibit smolt development and salinity tolerance and to delay smolt migration (Lerner et al., 2007; Madsen et al., 2004; McCormick et al., 2005).

Nonylphenol is a major degradation product of nonylphenol ethoxylate, which is a nonionic surfactant widely used in detergents, herbicides, pesticides, cosmetics, plastics, and spermicidal contraceptives. Wastewater treatment removes most nonylphenol ethoxylate and metabolites; however, nonylphenol is still commonly detected in rivers that receive industrial and municipal treatment effluents (Arukwe et al., 2000b; Naylor et al., 1992).

0166-445X/\$ – see front matter. Published by Elsevier B.V. http://dx.doi.org/10.1016/j.aquatox.2012.05.009

Concentrations as high as $325-1000 \mu g/L$ have been detected in receiving waters (Liber et al., 1999). Nonylphenol bioaccumulates in organisms and has been detected in amphipods, algae, fish, and other organisms (Ahel et al., 1993; Hecht et al., 2004; Liber et al., 1999).

Both estradiol and nonylphenol inhibit smolt development; treated fish display reduced gill Na⁺/K⁺-ATPase activity and expression, reduced numbers of gill chloride cells, and reduced seawater tolerance (Madsen et al., 1997; McCormick et al., 2005). Nonylphenol is an estrogen mimic; it competes with natural estrogens for binding to the estrogen receptor and induces expression of proteins involved in female egg production (vitellogenin and zona radiata proteins) (Arukwe et al., 2000a; Jobling and Sumpter, 1993; White et al., 1994). However, nonylphenol also appears to regulate some genes independently of the estrogen receptor (Larkin et al., 2002, 2003; Watanabe et al., 2004). Nonylphenol has weak affinity for the progesterone receptor in rats (Laws et al., 2000), activates pregnane X receptor-mediated transcription in cos-7 cells (Masuyama et al., 2000), and is a weak androgen receptor agonist in a heterologous system using a human receptor (Sohoni and Sumpter, 1998). In salmon, both nonylphenol and estradiol affect plasma concentrations of insulin-like growth factor I and thyroid hormones; nonylphenol also affects growth hormone levels (McCormick et al., 2005).

Smolting is an incredibly complex process that involves changes in appearance, behavior, and physiology. While many of the mechanisms that underlie these alterations are not well understood, it is well established that changes in gene expression are involved

^{*} Corresponding author. Tel.: +1 304 724 4579; fax: +1 304 724 4465.

E-mail addresses: lrobertson@usgs.gov (L.S. Robertson), smccormick@usgs.gov (S.D. McCormick).

in smolting. We used the GRASP 16K salmonid cDNA microarrays to identify genes whose expression was altered in Atlantic salmon smolts by exposure to nonylphenol. The expression of several genes identified by microarray analyses was verified using qRT-PCR. We then compared the results of this study with a companion study that identified genes differentially expressed in Atlantic salmon smolts compared to parr. We included a phenotypic examination of gill Na⁺/K⁺-ATPase activity, plasma thyroid hormones and cortisol, which have previously shown to be impacted by nonylphenol (McCormick et al., 2005), to verify the impacts of nonylphenol exposure.

2. Methods

2.1. Fish rearing and treatments

Atlantic salmon of Connecticut River stock obtained as parr from the White River National Fish Hatchery in fall 2005 were reared at the Conte Anadromous Fish Research Center in Turners Falls, MA. Fish were held in 1.5 m diameter fiberglass tanks receiving flow-through (4 L min⁻¹) Connecticut River water (Ca²⁺, 9.0 mg L⁻¹; Mg²⁺, 1.5 mg L⁻¹; Na⁺, 6.8 mg L⁻¹; K⁺, 1.1 mg L⁻¹; Cl⁻, 11.0 mg L^{-1} , pH 6.8–7.8) and maintained under natural photoperiod conditions. Fish were reared under ambient river temperatures through the fall and then at 10 $(\pm 1)^{\circ}$ C from February through the remainder of the study. Fish were fed to satiation twice daily with commercial feed (Zeigler Bros., Garners, PA, USA). In late January, fish were divided into two groups (parr and smolts) and reared in separate tanks. Fish smaller than 11 cm were considered parr and fish greater than 12 cm were considered smolts. Previous studies have established these size thresholds for smolt development in Atlantic salmon (McCormick et al., 2007).

For the nonylphenol exposure, ten smolts were injected with $150 \mu g/g$ body weight branched para-nonylphenol (CAS No. 84852-15-3) dissolved in vegetable oil and ten smolts were injected with vehicle control on day 0 and on day 4 as outlined in (McCormick et al., 2005); fish were sampled on day 7. Food was withheld throughout the treatment period, which occurred from April 17 to April 24.

On day 7, fish were euthanized in 200 mg L^{-1} MS-222 neutralized to pH 7.0; fork length to the nearest millimeter and weight to the nearest 0.1 g were recorded. Blood was drawn from the caudal vein into 1 mL ammonium heparinized syringes and spun at $8000 \times g$ for 5 min at 4 °C. Plasma was aliquoted and immediately stored at -80 °C for subsequent measurement of hormones. Four to six gill filaments were severed above the septum, placed in $100 \,\mu\text{L}$ of ice-cold SEI buffer (150 mM sucrose, 10 mM EDTA, 50 mM imidazole, pH 7.3) and immediately frozen at -80 °C. The remaining gill tissue was trimmed from gill arches for expression analysis. The gill, liver, pituitary, hypothalamus and both olfactory rosettes were collected in individual tubes, frozen immediately on dry ice and stored at -80 °C for microarray analyses.

2.2. Physiological assays

Na⁺/K⁺-ATPase activity was determined as described in (McCormick, 1993). Briefly, gill tissue was homogenized in 150 μ L of SEID (SEI buffer and 0.1% deoxycholic acid) and centrifuged at 5000 × g for 30 s. Two sets of duplicate 10 μ L samples were run, one set containing assay mixture and the other containing assay mixture and 0.5 mM ouabain. The resulting ouabain-sensitive ATPase activity (expressed as μ moles ADP/mg protein/h) was quantified via a kinetic assay run in 96-well microplates at 25 °C and read at a wavelength of 340 nm for 10 min. Protein concentrations were determined using a BCA (bicinchoninic acid) Protein Assay

(Pierce, Rockford, IL, USA). Both assays were run on a THERMOmax microplate reader using SOFTmax software (Molecular Devices, Menlo Park, CA, USA).

Plasma cortisol levels were measured by a validated direct competitive enzyme immunoassay as described by Munro and Stabenfeldt (Munro and Stabenfeldt, 1984) and modified by Carey and McCormick (Carey and McCormick, 1998). Assay range as defined by the standard curve was $1-200 \text{ ng mL}^{-1}$. The lower detection limit was 0.3 ng mL^{-1} . Based on a pooled plasma sample that was included on each plate, the average intra-assay variation was 5.5% (n=10) and the average inter-assay variation was 8.8% (n=10). Thyroxine (T_4) and 3,5,3'-triiodo-L-thyronine (T_3) concentrations were measured by a direct radioimmunoassay described by Dickhoff et al. (1978) and modified by McCormick et al. (1995). The assay range as defined by the standard curve was $1-64 \text{ ng mL}^{-1}$ for T_4 and $0.5-16 \text{ ng mL}^{-1}$ for T_3 . Intra- and inter-assay coefficients of variation for these assays were 4-11% and 3-5%, respectively.

2.3. RNA isolation

Total RNA was extracted from tissue samples frozen at -80 °C using RNeasy kits (Qiagen, Inc., Valencia, CA, USA) according to manufacturer instructions; all samples were treated with RNase-free DNase (Qiagen, Inc., Valencia, CA, USA) to remove genomic DNA. RNA quantity was measured with an ND-1000 UV–Vis Spectrophotometer (Nanodrop Technologies/Thermo-Fisher Scientific, Inc., Wilmington, DE, USA). Quality of the RNA was assessed using a 2100 Bioanalyzer (Agilent Technologies, Inc., Santa Clara, CA, USA). One sample used in the microarray hybridizations had an RNA Integrity Number (RIN) (Schroeder et al., 2006) of 8.8; the other 79 samples used in these hybridizations had a RIN number greater than or equal to 9.0. RNA samples were stored at -80 °C.

2.4. Microarray hybridization

GRASP 16K cDNA microarrays spotted with 16,006 features were purchased from the Genomics Research on All Salmon consortium (GRASP, http://web.uvic.ca/grasp/). These microarrays have been validated for salmonid species (von Schalburg et al., 2005). Slide pre-processing, sample hybridization, and post-hybridization washing were performed as recommended by GRASP.

The pituitary is a very small organ and only a small amount of total RNA was extracted (1–3 µg total RNA) from each pituitary. Due to the small amount of total RNA extracted, pituitary RNA was amplified in a single round and then labeled using the Ambion Amino Allyl MessageAmpII aRNA amplification kit (Applied Biosystems/Life Technologies Corporation, Carlsbad, CA, USA) according to manufacturer's instructions. Hypothalamus, liver, gill, and olfactory rosette RNA samples were not amplified prior to labeling. These samples were reverse transcribed and labeled using the Superscript Plus Indirect cDNA Labeling System (Invitrogen/Life Technologies Corporation, Carlsbad, CA, USA). All samples were labeled with Alexa Fluor 555 or 647 dyes (Invitrogen/Life Technologies Corporation, Carlsbad, CA, USA). Qiaquick PCR Purification spin columns (Qiagen, Inc., Valencia, CA, USA) were used to clean the cDNA following first strand synthesis and following labeling with Alexa Fluor dves.

Microarray slides from the same print batch were used for each tissue comparison. Ambion SlideHyb no. 2 (Applied Biosystems/Life Technologies Corporation, Carlsbad, CA, USA) was used for hybridization. The slides were incubated at 50 °C, shaking slowly, for 16–18 h. After hybridization, slides were washed, dried by centrifugation, and then immediately scanned at 10 µm resolution in a ScanArray Gx microarray scanner (Perkin Elmer, Waltham, MA, USA). Laser power and PMT voltage were adjusted using the Line Scan Protocol (ScanArray Microarray Analysis System version

L.S. Robertson, S.D. McCormick / Aquatic Toxicology 122-123 (2012) 36-43

2	0
- 3	ð

Table 1

Primer sequences used for real-time qRT-PCR.

Gene or best BLAST hit ^a	Accession	Forward primer	Reverse primer
Apolipoprotein A-IV ^{a,f}	CB505965	GATCTGAAGGCCCAGCTTACTG	GGTTTGGGCCATTTAGTTGGT
Cathepsin D ^a	CB504794	CGGAGCAACCCCACTCAT	GCATGGTGGGAATCTTATCACA
EF1A _A	AF321836	CCCCTCCAGGACGTTTACAAA ^b	CACACGGCCCACAGGTACA ^b
EF1A _B	BG933853	TGCCCCTCCAGGATGTCTACb	CACGGCCCACAGGTACTG ^b
Endoplasmin ^{a, f}	CA059329	CCCAACAGCCACCAATCAA	TTGGGCAAGAGCACAAACAG
EST: ssallna016029 ^e	CA769178	GGCTGTGCTAGGCTGGAGTT	CGGTTCTGATGGCAGCACTT
Ferritin heavy polypeptide 1–1	NM_001139722 ^d	GGGTGCCCCCAGAAC	CTTCCCCAAAGTGTGTTTGTCA
Ferritin middle subunit	NM_001123658	TGCACAAGATTGCCTCTGACA	AGGTAATGGGTCTCCAGGAAGTC
Ferritin lower subunit (fri3)	NM_001141424	CCCGATGAGCAGCAGGTACT	CCACATACCTGATTGGCAGATG
High-affinity copper uptake protein 1ª	CA042004	TGAAGATCGGCAGGGAGTTC	TGGAGTTATAGCGCACGTTGA
IGF-I	NM_001123623	CAGTTCACGGCGGTCACATA	GCTCGCAACTCTGGAAGCA
Polyubiquitin-C ^a	CA055128	TCCTCTGCTCTTAACGATGTCTTATC	GGTCAATGGGTGGGATTGG
Proteasome subunit beta type 3ª	BU965861	TCCCAAGCCATGCTGAATG	AATGACACCCATGCCTGAAAC
40S Ribosomal protein S20 ^c	BG936672	GCAGACCTTATCCGTGGAGCTA ^b	TGGTGATGCGCAGAGTCTTG ^b
Stress-associated endoplasmic reticulum protein 1 ^{a,f}	CA044731	GGATGGGCATGTAGGATGACA	TGGTTGGGCGGATACTGAGT
Tubulin alpha chain ^a	CK990304	GGCTTCAAGGTGGGTATCAACT	GCCAGGTCTCCTCCAGGAA
Vitellogenin ^{a,f}	CB515390	TCCCAGTGAGGACCACCTACA	CAGAGCGGTTCAGGTTGGA
Zona pellucida sperm-binding protein 3ª	CK991165	TGTCTGATCGATGCCAAAGTG	TGCGGAACGAGGCATGA
^a Best BLAST hit, see methods for BLAST parameters.			

^b Primers designed by Olsvik et al. (Olsvik P.A., Lie K.K., Jordal A.E., Nilsen T.O., Hordvik I., 2005. Evaluation of potential reference genes in real-time RT-PCR studies of Atlantic salmon. BMC Mol Biol 6,21).

^c Reference gene for qPCR.

^d Ferritin heavy polypeptide 1–1 primers match both NM.001139722 (ferritin, heavy polypeptide 1–1) and NM.001123657 (ferritin, heavy polypeptide 1–2).

^e No BLAST hit with *e*-value <1e-3.

^f BLAST blastx query coverage less than 50%.

3.0, Perkin Elmer) such that the signal from the Alexa Fluor 555 dye was approximately equal to the signal from the Alexa Fluor 647 dye for each slide and the signal was saturated for approximately 0.3–0.5% of the features on the slide.

2.5. Array experimental design

Forty microarray slides were used to compare expression in nonylphenol-injected fish with expression in vehicle-injected fish: eight two-channel arrays each for liver, gill, olfactory rosettes, hypothalamus, and pituitary tissues. RNA was isolated from the tissues of eight nonylphenol-injected fish (six males and two females) and eight vehicle-injected fish (two males and six females) and reverse transcribed separately (samples were not pooled); each slide represented a biological replicate. For each tissue, the eight arrays were balanced for dye: nonylphenol-injected fish were labeled with Alexa Fluor 555 and vehicle-injected fish were labeled with Alexa Fluor 647 on four slides, nonylphenol-injected fish were labeled with Alexa Fluor 555 on four slides. Liver, gill, hypothalamus, pituitary, and olfactory rosette tissues were analyzed separately for each experiment, resulting in five distinct analyses.

2.6. Microarray analysis

Spot-finding was done with ScanArray Express Microarray Analysis System software version 4.0 using the adaptive circle method. Spot-finding images were manually reviewed and improperly identified spots were manually adjusted.

The TM4 Microarray Software Suite was used for pre-processing, filtering, normalization, and statistical analysis of the microarray data (Saeed et al., 2003). Files containing median spot intensities and median background intensities were imported into the TM4 Microarray Software using Express Converter version 2.1 utility; median background intensities were subtracted from median spot intensities. Converted intensity files were imported into TM4 Midas version 2.20 with RMA and spots where the backgroundcorrected intensity was less than zero for either channel (dye) were removed. The log₂ intensity ratios for each array were normalized using a local Lowess normalization (by subarray or block) and standard deviation regularization for the blocks. Lowess correction was done by block, or pin group, to correct for any differences between spotting pin or spatial differences across the slide. One-class Significance Analysis of Microarrays (SAM; TM4 MeV version 4.2) was used to identify genes with significantly different expression between conditions; the delta value was minimized such that the median false discovery rate (FDR) was less than 1%. Genes with significantly different expression between conditions (nonylphenol versus vehicle control) were also identified by t-test (TM4 MeV version 4.2). Three different criteria were used to identify potentially significant differentially expressed genes: p-value < 0.001, p < 0.01 and fold change \geq 1.20, *p* < 0.01 and fold change \geq 2.0. To investigate false discovery rate with *t*-test analyses, *q*-values were calculated using the QVALUE software written by Alan Dabney and John Storey (Storey and Tibshirani, 2003).

Table 2

Number of significantly differentially expressed features identified by *t*-test with p < 0.001 (*t*-test), by *t*-test with p < 0.01 and at least two-fold change in expression (*t*-test and $2 \times$), and by SAM analysis (SAM) in nonylphenol-treated fish compared to control fish, by tissue.

	<i>t</i> -Test, <i>p</i> < 0.001	<i>t</i> -Test, <i>p</i> < 0.01			SAM
		Total	\geq 1.20-Fold change	≥2.00-Fold change	
Liver	175	678	459	84	124
Gill	1	27	7	0	0
Hypothalamus	8	91	23	6	0
Pituitary	1	47	24	1	0
Rosettes	7	57	13	2	0

L.S. Robertson, S.D. McCormick / Aquatic Toxicology 122-123 (2012) 36-43

The translated sequences for features identified as differentially expressed in microarray analyses were compared to the SwissProt database using the blastx program (http://blast.ncbi.nlm.nih.gov). The name of the blast hit having the lowest *e*-value is given as the "best BLAST hit" for each feature (Tables 1, 3 and 4). Features that did not have a BLAST hit with *e*-value less than 1e–3 and features where the region of sequence similarity covered less than 50% of the translated feature sequence are indicated. Note that the best BLAST hit listed in these tables was not always the same as the annotation provided by the GRASP consortium.

2.7. Real-time PCR analysis

Real-time PCR was used to validate the expression of select genes identified in microarray analyses (Rajeevan et al., 2001). Gene expression was measured in twenty fish (ten fish exposed to nonylphenol, ten fish exposed to vehicle control), including the sixteen fish used for the microarray hybridizations. Total RNA was reverse transcribed using the High Capacity cDNA Reverse Transcription Kit (Applied Biosystems/Life Technologies Corporation, Carlsbad, CA, USA). Real-time gRT-PCR reactions were done in triplicate on an ABI7900HT using Power SYBR Green Master Mix (Applied Biosystems/Life Technologies Corporation, Carlsbad, CA, USA). Primers used in real-time PCR are listed in Table 1. Mean expression of target gene divided by mean expression of a reference gene was reported. Ribosomal protein S20 was used as a reference gene. Statistical significance of real-time PCR comparisons between control and nonylphenol-treated fish were examined by Mann–Whitney U-test.

3. Results

3.1. Physiological effects of nonylphenol exposure

Gill Na⁺/K⁺-ATPase activity was reduced by approximately 25% after one week exposure to 150 μ g/g nonylphenol (Fig. 1, *p* = 0.004). Plasma cortisol was reduced by 60% in nonylphenol treated fish (Fig. 1, *p* = 0.019, Mann–Whitney *U*-test). Plasma T₄ was not significantly affected by nonylphenol treatment, whereas plasma T₃ was reduced by 20% (Fig. 1, *p* = 0.03).

3.2. Differentially expressed genes identified by microarray analyses

GRASP 16K microarrays were used to identify genes in the gill, liver, olfactory rosettes, hypothalamus, and pituitary that were differentially expressed between nonylphenol-treated fish and control fish. The microarray data discussed here have been deposited in NCBI's Gene Expression Omnibus (Edgar et al., 2002) and are accessible through GEO Series accession number GSE33715 (http://www.ncbi.nlm.nih.gov/geo/query/acc.cgi?acc=GSE33715). The number of differentially expressed features identified by Significance Analysis of Microarrays (SAM) or by *t*-test at two significance levels (p < 0.001 and p < 0.01) for all five tissues is listed in Table 2. Table 3 lists the features identified by SAM.

By far the largest effect of nonylphenol on gene expression was seen in the liver (Table 2), with little to no effect seen in gills, olfactory rosettes, hypothalamus, or pituitary. 124 features were significantly different in the livers of nonylphenol treated fish compared to control fish by SAM, with a median false discovery rate less than 1%; all 124 features are more abundant in nonylphenol-treated fish than control fish.

These microarray data were also analyzed using *t*-tests; *q*-values were calculated for the entire dataset to investigate the false discovery rate (Storey and Tibshirani, 2003). Using a p < 0.01 significance level for the *t*-test, 678 features were significantly different

Table 3

Features significantly (SAM) upregulated in the liver of juvenile Atlantic salmon treated with nonylphenol compared to control fish.

Best BLAST hit	Accession	Fold change
28S ribosomal protein S17 ^c	CB504359	3.8
40S Ribosomal protein S13	CB491393	6.8
60S ribosomal protein L39 ^c	CA038333	1.2
ADP-ribosylation factor 4	CB499782	1.6
Apolipoprotein A-IV	CB205965	2.1
Apolipoprotein C-I	CB495958 CB491411	6.4
ATP synthase lipid-binding protein, mitochondrial	CB493846	1.1
Bone morphogenetic protein 4	CA056395	1.3
Calmodulin-A	CA050185	1.4
Cytochrome c oxidase assembly factor 5 ^c	CB486148	2.3
Endoplasmin ^c	CA059329	1.5
Endoplasmin ^c	CB492804	1.5
EST: omykrbn2502215 ^b	CB495547 CB496826	1.9
EST: omykrbna502358 ^b	CB497380	1.4
EST: omykrbna503004 ^b	CB497817	2.0
EST: omykrbna503022 ^b	CB497857	1.8
EST: omykrbna503044 ^b	CB497906	1.5
EST: omykrbna504040 ^b	CB494172	1.5
EST: omykrbna509054 ^b	CB492399	2.5
EST: omykrtcl003081	CB491/21	/.6
EST: ssalmae50/351 ^b	CR505701	1.8
FST: ssalmge505245 ^b	CB503310	3.0
EST: ssalnwh501337 ^b	CB510155	1.6
EST: ssalnwh505093 ^b	CB509722	4.1
EST: ssalnwh505328 ^b	CB509927	1.7
EST: ssalob004012 ^b	CA036615	1.3
EST: ssalob502029 ^b	CB508626	1.6
EST: ssalob504369 ^b	CB508658	3.3
EST: ssalpinb501118 ^b	CA042416	2.7
EST: scalphp548221 ^b	CR514261	3.2
EST: ssalrga501057 ^b	CA051359	11
EST: ssalrga508035 ^b	CA060454	1.8
EST: ssalrga508120 ^b	CA060809	1.6
EST: ssalrga516287 ^b	CA061210	3.4
EST: ssalrgb503074 ^b	CA054826	1.1
EST: ssalrgb515125 ^b	CA052877	2.5
EST: scaligD536181 ^b	CA769251	1.4
EST: ssalrgb536182 ^b	CA056442	1.3
EST: ssalrgb536204 ^b	CA056568	1.4
EST: ssalrgb536250 ^b	CA056640	1.3
EST: ssalrgb536253 ^b	CA056644	2.0
EST: ssalrgb536254 ^b	CA056647	2.3
EST: ssalrgb536273 ^b	CA056591	1.6
EST: scaligb536217 ^b	CA056038	1.5
EST: ssalrgb536323 ^b	CA056958	1.5
EST: ssalrgb536348 ^b	CA057016	1.3
EST: ssalrgb551343 ^b	CB512669	2.0
EST: ssalrgb555204 ^b	CB516919	1.8
EST: ssalrgb555265 ^b	CB517012	1.1
EST: ssalshc502278 ^b	CA040841	2.5
EST: ssalshc503347 ^b	CA040291	1.4
EST: SSAISTKC002051 ² Fatty acid_binding protein_beart	CR050094 CR489347	2.0
Ferritin, middle subunit	CB487639	8.1
Ferritin, middle subunit	CB506201	6.9
Ferritin, middle subunit	CB505282	6.3
Ferritin, middle subunit	CB503780	5.6
Ferritin, middle subunit	CN442541	5.3
Ferritin, middle subunit ^c	CA042437	4.9
Ferritin, middle subunit	CB502663	4.7
Ferritin middle subunit	CB490370 CB507396	4.0 4.5
Ferritin, middle subunit	CB497884	4.4
Ferritin, middle subunit ^c	CA768211	4.3
Ferritin, middle subunit	CK991031	4.2
Ferritin, middle subunit	CB510731	3.7
Ferritin, middle subunit ^c	CA037206 ^a	3.7
Ferritin, middle subunit	CA045790	3.3

Author's personal copy

L.S. Robertson, S.D. McCormick / Aquatic Toxicology 122-123 (2012) 36-43

Table 3 (Continued)

Best BLAST hit	Accession	Fold
		change
Ferritin middle subunit	CB508292	2.4
Ferritin, middle subunit	CA056696	2.3
Ferritin, middle subunit ^c	CK990667	2.2
Ferritin, middle subunit	CA056744	1.8
Ferritin, middle subunit	CK990310	1.6
Ferritin, middle subunit	CA052539	1.5
Ferritin, middle subunit	CK990827	1.5
Galectin-3 ^c	CB514642	1.2
Heat shock protein HSP 90-alpha 1	CB498021	1.7
Heat shock protein HSP 90-beta	CA039749	1.7
High affinity copper uptake protein 1	CA042004	2.5
Keratinocyte-associated protein 2 ^c	CB506143	1.3
KIF1-binding protein homolog ^c	CB496780	5.4
Mitotic-spindle organizing protein 2 ^c	CB515170	1.3
Nuclear receptor subfamily 0 group B member 2	CA037720	2.2
PCNA-associated factor	CB490001	1.7
Polyubiquitin-B	CA055128	1.6
Polyubiquitin-B	CB518072	1.1
Proteasome subunit alpha type-1 ^c	CB494281	1.8
Proteasome subunit alpha type-2	CB498208	2.2
Proteasome subunit alpha type-3	CB496456	2.6
Proteasome subunit alpha type-4	CB487801	3.2
Proteasome subunit alpha type-6	CA038117	2.1
Proteasome subunit beta type-3	CB493627	2.0
Proteasome subunit beta type-3	BU965861	1.4
Protein FAM71B	CB516929	1.5
Protein transport protein Sec61 subunit beta ^c	CA044589	1.5
Protein transport protein Sec61 subunit gamma ^c	CB497886	1.2
Small ubiquitin-related modifier 3 ^c	CB490926	3.0
Stathmin	CB486047	2.2
Stress-associated endoplasmic reticulum protein 1 ^c	CA044731	1.9
Stress-associated endoplasmic reticulum protein 1 ^c	CB512613	1.6
Synapse differentiation-inducing gene protein 1-like ^c	CA056593	1.5
T-complex protein 1 subunit gamma	CA056693	1.4
Thioredoxin	CA041451	1.5
TSC22 domain family protein 3 ^c	CB509522	2.1
Tubulin alpha chain	CB494048	3.2
Tubulin alpha chain	CB493545	3.0
Tubulin alpha chain ^c	CB511276	3.0
Tubulin alpha chain ^c	CB497308	2.5
Tubulin alpha chain Tubulin alpha abaint	CK990304	2.4
Tubulin alpha chain ^c	CA047720	1.8
Tubulin alpha chain ^c	CA046426	1./
Tubulin alpha chain	CAU58890	1.0
Tubulin dipila Cildili Tubulin bota chain	CA056240	1.1
Tubulin beta 1 chain	CR511172	1.9
Vitellogenin	CA051622	2.0
Vitellogonin ^c	CR515200	4./
Vicinogenini ⁻	CE013390	5./ 22
2014 penuciua spenii-pinuliig protein 3	CR504250	2.5
40S Pibosomal protein S12	CD304339	5.0 6.9
405 KIDOSOIIIAI PROLEIII 513	CB491393	U.ð

^a Accession number for feature identified in microarray analysis; see Table 1 for accession number of sequence used to design qPCR primers.

^b No BLAST hit with *e*-value <1e–3.

^c Query coverage less than 50%.

between nonylphenol-treated and control fish; the largest *q*-value in this set was 0.169. Using a *q*-value threshold of 0.05, 146 genes are significant. The overall proportion of true null hypotheses in this dataset, π_0 , was 0.73.

3.3. Validation of microarray results by qRT-PCR

qRT-PCR was used to measure the mRNA abundance of ten genes identified by SAM as differentially expressed in the liver of nonylphenol-treated fish compared to control fish (Table 4). These ten genes include the two genes most frequently identified (ferritin middle subunit and tubulin alpha chain); the gene with the highest observed fold-change (ferritin); and a gene known to be affected by nonylphenol (vitellogenin). These genes were

also selected to represent a range of fold change in expression, including several genes with observed fold changes above 2.0 (high affinity copper uptake protein 1 - 2.5×, zona pellucida spermbinding protein 3 – 2.3, apolipoprotein A-IV – $2.1\times$) and several genes with observed fold changes below 2.0 (stress-associated endoplasmic reticulum protein 1 – 1.9×, polyubiquitin-B – 1.6×, endoplasmin – $1.5 \times$, proteasome subunit beta type-3 – $1.4 \times$). Eight of the ten genes were significantly differentially expressed between nonylphenol-treated fish and control fish, by qRT-PCR. The abundances of apolipoprotein A-IV and high-affinity copper uptake protein 1 were not significantly different between nonylphenoltreated fish and control fish, by qRT-PCR. Cathepsin D (Genbank accession number CB504794) barely missed the criteria used for t-test (p=0.001095, 1.78-fold change) and was not identified by SAM; expression of cathepsin D was significantly different between nonylphenol-treated fish and control fish by qPCR.

Abundance of the ferritin middle subunit (annotated as ferritin heavy chain on the GRASP array) was significantly higher in the liver of nonylphenol-treated fish than in control fish by microarray analysis (22 features identified by SAM; 1.5-fold to 8.1-fold higher in nonylphenol-treated fish). The RNA abundance of all three ferritin subunits was measured by qRT-PCR. Expression of the ferritin middle subunit was 18.9-fold higher in nonylphenol-treated fish (p < 0.0001) and the ferritin heavy subunit was 1.7-fold higher in nonylphenol-treated fish (p = 0.0007), but the expression of the ferritin light subunit was decreased (fold change 0.36, p = 0.0355).

Insulin-like growth factor 1 (IGF-1) was decreased in the liver of nonylphenol-treated fish compared to control fish by microarray analysis (fold change = 0.60, *t*-test, p < 0.01). The decrease in IGF-I mRNA abundance in the liver of nonylphenol-treated fish was greater by qRT-PCR (fold change = 0.11, p = 0.0002).

Elongation factor genes EF1A_A and EF1A_B, 40S ribosomal protein S20, and the EST: ssallna016029 were considered as possible reference genes. Expression of both EF1A_A and EF1A_B was increased slightly, but significantly, in the liver of nonylphenol-treated fish compared to control fish (data not shown). Expression of 40S ribosomal protein S20 and the EST: ssallna016029 was not significantly different between nonylphenol-treated fish and control fish. The 40S ribosomal protein S20 was used for normalization of qRT-PCR assays.

4. Discussion

4.1. The effect of nonylphenol on gene expression in Atlantic salmon smolts

We used the GRASP 16K microarray to identify transcriptional effects of nonylphenol in the gill, liver, olfactory rosettes, pituitary, and hypothalamus. The liver was the most responsive in terms of number of genes transcriptionally altered by exposure to nonylphenol. The effect of nonylphenol on transcription outside the liver was small. Genes involved in protein biosynthesis, protein folding, and protein transport; proteolysis, protein catabolism, and ubiquitindependent protein catabolism; nucleosome assembly, cell cycle, cell differentiation, microtubule-based movement (tubulin subunits), electron transport, and response to stress were increased in the liver of nonylphenol-treated fish. Expression of vitellogenin (Genbank accession numbers CA051633 and CB515390), a commonly used biomarker for exposure to estrogenic compounds, was increased in the liver of nonylphenol-treated fish, as was expression of zona pellucida sperm-binding protein 3 (Genbank accession numbers CK991165 and CB490454).

The largest fold-change in response to nonylphenol and the most commonly identified feature was the middle subunit of ferritin. In teleost fish, a heavy subunit and a middle subunit for ferritin

L.S. Robertson, S.D. McCormick / Aquatic Toxicology 122-123 (2012) 36-43



Fig. 1. Physiological and endocrine differences between Atlantic salmon smolts treated with vehicle control and nonylphenol (150 g/g) for one week, and sampled on day 7. Values are mean + standard error of 10–12 fish per group. Asterisk indicates a significant difference between part and smolt (*p* < 0.05, *t*-test).

have been identified (Andersen et al., 1995; Liu et al., 2010; Zhang et al., 2010); a putative cDNA for ferritin lower subunit has also been identified (Leong et al., 2010). Nonylphenol exposure significantly increased expression of the ferritin middle subunit in Atlantic salmon smolts, as determined by our microarray analysis. Expression of both the ferritin heavy and middle subunits was increased by nonylphenol exposure, but expression of the ferritin lower subunit was decreased (qRT-PCR). Expression of the ferritin heavy subunit is not significantly different between parr and smolts (Robertson and McCormick, unpublished results). Ferritin is an iron storage protein that is involved in iron homeostasis and response to infection. Expression of the ferritin heavy subunit increases upon infection in channel catfish (Peatman et al., 2007) and in sea bass (Neves et al., 2009); this increase in ferritin expression may result in the sequestration of available iron, thus starving the pathogen of a required nutrient. The ferritin middle subunit from tongue sole displays antimicrobial activity (Wang et al., 2011). Ferritin also appears to be involved in reproductive development. The ferritin heavy chain is one of the genes most significantly upregulated by estradiol and by nonylphenol in male zebrafish liver (Ruggeri

et al., 2008). Ferritin expression also increases in the ovaries of coho salmon during oogenesis (Luckenbach et al., 2008).

In this study, gill Na⁺/K⁺-ATPase activity was reduced by approximately 25% in nonylphenol-treated fish compared to control fish. The salmon gill Na⁺/K⁺-ATPase (NKA) enzyme consists of two subunits (α and β): the alpha subunit contains the catalytic sites and the beta subunit is involved in protein maturation and insertion into the membrane. Salmon have multiple isoforms for each subunit; two of the NKA- α 1 isoforms are differentially expressed during smolting and seawater transfer (Nilsen et al., 2007). The GRASP 16K array has six features representing the alpha subunit and four features representing the beta subunit. None of these features were identified as differentially expressed in nonylphenoltreated fish compared to control fish in any tissue.

Due to an unintended imbalance in the sex ratio between nonylphenol-treated fish (two females and eight males) and control fish (eight females and two males), we were unable to assess sexspecific changes in gene expression. Nonylphenol has been shown to increase expression of luteinizing hormone beta in the pituitary of females, but not males, and to increase the expression of the

Table 4

Expression of selected genes as measured by qRT-PCR in the livers of Atlantic salmon. Fold change is the mean normalized abundance in nonylphenol-treated fish divided by the mean normalized mRNA abundance in control fish; Mann–Whitney test *p*-value.

Gene (or best BLAST hit ^a)	Accession	Fold change	<i>p</i> -value
Apolipoprotein A-IV ^a	CB505965	2.86	0.9698
Endoplasmin ^{a,b}	CA059329	3.26	0.0029
Ferritin middle subunit	NM_001123658	17.38	< 0.0001
High-affinity copper uptake protein 1 ^a	CA042004	3.97	0.2567
Polyubiquitin	CA055128	2.88	< 0.0001
Proteasome subunit beta type 3ª	BU965861	4.26	< 0.0001
Stress-associated endoplasmic reticulum protein 1 ^a	CA044731	3.40	< 0.0001
Tubulin alpha chain ^a	CK990304	10.12	< 0.0001
Vitellogenin ^a	CB515390	16,796.54	< 0.0001
Zona pellucida sperm-binding protein 3ª	CK991165	31.11	< 0.0001
Cathepsin D ^a	CB504794	3.76	< 0.0001

^a Best BLAST hit, see methods for BLAST parameters.

^b GenBank accession no. CB493461 identified in microarray analyses; GenBank accession no. CA059329 used to design primers for qPCR.

transcription factor Pit-1 in the pituitary of males, but not females (Yadetie and Male, 2002). However, previous results indicate that endocrine disruption of smolt development occurs similarly in both sexes (31).

4.2. Nonylphenol treatment affected expression of very few genes known to be regulated by smolting

Robertson and McCormick (unpublished results) used the GRASP 16K cDNA arrays to identify differentially expressed genes in Atlantic salmon smolts compared to parr in the liver, gill, hypothalamus, pituitary, and olfactory rosettes at a single timepoint. The microarray data for this parr-smolt comparison have been deposited in NCBI's Gene Expression Omnibus (Edgar et al., 2002) and are accessible through GEO Series accession number GSE33711 (http://www.ncbi.nlm.nih.gov/geo/query/acc.cgi?acc=GSE33711). There was no overlap between the microarray features significantly (SAM) affected by nonylphenol (this study) and microarray features significantly affected in smolts compared to parr (Robertson and McCormick, unpublished results). However using a less-stringent analysis (t-test p<0.01, filtered by fold-change threshold of 1.2fold), ten features were identified as altered both by treatment with nonylphenol and also by smolting. Interestingly, all ten features changed in the opposite direction in nonylphenol-treated fish compared to vehicle than in smolts compared to parr, as was the case with IGF-I. The transcripts for seven genes were more abundant in the livers of smolts compared to parr, but less abundant in the livers of nonylphenol-treated smolts compared to control smolts: complement C3, glutathione S-transferase theta 1, pancreatic progenitor cell differentiation and proliferation factor B (PPDPFB), secreted phosphoprotein 24, and three sequences of unknown function (Genbank accession numbers CB511254, CB510516, and CA058127). The transcripts for three genes were less abundant in the livers of smolts compared to parr, but more abundant in the livers of nonylphenol-treated smolts compared to control smolts: glutathione S-transferase P, PCNA-associated factor, and one sequence of unknown function (Genbank accession number CA041709). The sequence CA041709 has no blastx hit with *e*-value <1e-3 in the non-redundant SwissProt database; however, a tblastx search finds several nucleotide sequences predicted to encode ependymin or ependymin-related proteins with similarity to the translated sequence for CA041709. In general, higher fold changes in expression were seen with the short exposure to nonylphenol compared to the fold changes seen at a single time point in the middle of smolting, a developmental process that takes several months.

Our analysis of changes in gill Na⁺/K⁺-ATPase activity and plasma hormone levels clearly indicated that the nonylphenol treatment we employed had a negative effect on smolt development, so we were surprised by the relatively small number of genes (particularly regulatory genes) related to smolt development that were also affected by nonylphenol. It should be noted, however, that relatively few genes involved in the endocrine control of smolt development were identified in our microarray approach (Robertson and McCormick, unpublished results). This may be due to the fact that changes in hormone gene transcription during smolting occur over a period of weeks and are relatively small in magnitude. In addition, substantial regulation may occur through non-transcriptional or through indirect pathways. Recent studies in coho salmon indicate neither pituitary transcription or circulating levels of thyroid stimulating hormone (TSH) are altered during smolt development (Larsen et al., 2011), indicating that increases in circulating levels of thyroid hormones during smolting are indirect, possibly through sensitivity to TSH. Our inability to detect substantial number of regulatory genes affected by nonylphenol may in part be due to the gradual nature and complexity of smolt development. In addition, only certain aspects of smolt development, such as osmoregulation and migration, may be strongly impacted by estrogenic compounds.

We were able to detect dramatic decreases in IGF-I mRNA abundance in response to nonylphenol treatment. This was predicted based on previous studies indicating that circulating levels of IGF-I decrease in response to nonylphenol (McCormick et al., 1995). The increase in salinity tolerance and gill Na⁺/K⁺-ATPase activity that occurs during smolting is in part supported by increases in plasma levels and local production of IGF-I (McCormick, 2001). Therefore, the decrease in gill Na⁺/K⁺-ATPase activity following nonylphenol exposure in the present study may be driven by decreased IGF-I transcription. Since cortisol is also involved in upregulation of gill Na⁺/K⁺-ATPase (McCormick, 2001), it seems likely that the observed decrease in cortisol is also involved.

The widely used surfactant nonylphenol is an endocrine disrupting compound that has been implicated in historical declines of Atlantic salmon. In addition, nonylphenol is known to inhibit smolting and delay smolt migration. Using a genome-wide approach, we investigated the effect of nonylphenol on gene expression in several tissues of Atlantic salmon smolts; the most dramatic effects were found in the liver. Through comparison with a companion study that identified genes that are differentially expressed in smolts compared to parr we found ten genes whose mRNA abundance was affected in the opposite direction by nonylphenol compared to smolt development, and these are of particular interest for further study.

Acknowledgements

The excellent technical assistance of Jamie Marie Marranca, Amy Regish and Michael O'Dea are gratefully acknowledged. This work was funded by the U.S. Geological Survey. Any use of trade, product, or firm names is for descriptive purposes only and does not imply endorsement by the U.S. Government.

References

- Ahel, M., McEvoy, J., Giger, W., 1993. Bioaccumulation of the lipophilic metabolites of nonionic surfactants in fresh-water organisms. Environmental Pollution 79, 243–248.
- Andersen, O., Dehli, A., Standal, H., Giskegjerde, T.A., Karstensen, R., Rorvik, K.A., 1995. 2 ferritin subunits of Atlantic Salmon (*Salmo-Salar*) – cloning of the liver Cdnas and antibody preparation. Molecular Marine Biology and Biotechnology 4, 164–170.
- Arukwe, A., Celius, T., Walther, B.T., Goksoyr, A., 2000a. Effects of xenoestrogen treatment on zona radiata protein and vitellogenin expression in Atlantic salmon (Salmo salar). Aquatic Toxicology 49, 159–170.
- Arukwe, A., Thibaut, R., Ingebrigtsen, K., Celius, T., Goksoyr, A., Cravedi, J.-P., 2000b. In vivo and in vitro metabolism and organ distribution of nonylphenol in Atlantic salmon (*Salmo salar*). Aquatic Toxicology 49, 289–304.
- Carey, J.B., McCormick, S.D., 1998. Atlantic salmon smolts are more responsive to an acute handling and confinement stress than parr. Aquaculture 168, 237–253.
- Dickhoff, W.W., Folmar, L.C., Gorbman, A., 1978. Changes in plasma thyroxine during smoltification of coho salmon, *Oncorhynchus kisutch*. General and Comparative Endocrinology 36, 229–232.
- Edgar, R., Domrachev, M., Lash, A.E., 2002. Gene expression omnibus: NCBI gene expression and hybridization array data repository. Nucleic Acids Research 30, 207–210.
- Fairchild, W.L., Swansburg, E.O., Arsenault, J.T., Brown, S.B., 1999. Does an association between pesticide use and subsequent declines in catch of Atlantic salmon (*Salmo salar*) represent a case of endocrine disruption? Environmental Health Perspectives 107, 349–358.
- Gephard, S., McMenemy, J., 2004. An overview of the program to restore Atlantic salmon and other diadromous fishes to the Connecticut River with notes on the current status of these species in the river. Connecticut River Ecological Study (1965–1973) Revisited: Ecology of the Lower Connecticut River 1973–2003 9, 287–317.
- Hecht, S.A., Gunnarsson, J.S., Boese, B.L., Lamberson, J.O., Schaffner, C., Giger, W., Jepson, P.C., 2004. Influences of sedimentary organic matter quality on the bioaccumulation of 4-nonylphenol by estuarine amphipods. Environmental Toxicology and Chemistry 23, 865–873.

L.S. Robertson, S.D. McCormick / Aquatic Toxicology 122-123 (2012) 36-43

- Jobling, S., Sumpter, J.P., 1993. Detergent components in sewage effluent are weakly oestrogenic to fish: an in vitro study using rainbow trout (*Oncorhynchus mykiss*) hepatocytes. Aquatic Toxicology 27, 361–372.
- Larkin, P., Sabo-Attwood, T., Kelso, J., Denslow, N.D., 2002. Gene expression analysis of largemouth bass exposed to estradiol, nonylphenol, and p.p'-DDE. Comparative Biochemistry and Physiology Part B Biochemistry and Molecular Biology 133B, 543–557.
- Larkin, P., Sabo-Attwood, T., Kelso, J., Denslow, N.D., 2003. Analysis of gene expression profiles in largemouth bass exposed to 17-beta-estradiol and to anthropogenic contaminants that behave as estrogens. Ecotoxicology 12, 463–468.
- Larsen, D.A., Swanson, P., Dickhoff, W.W., 2011. The pituitary-thyroid axis during the parr-smolt transformation of Coho salmon, *Oncorhynchus kisutch*: quantification of TSH beta mRNA, TSH, and thyroid hormones. General and Comparative Endocrinology 171, 367–372.
- Laws, S.C., Carey, S.A., Ferrell, J.M., Bodman, G.J., Cooper, R.L., 2000. Estrogenic activity of octylphenol, nonylphenol, bisphenol A and methoxychlor in rats. Toxicological Sciences 54, 154–167.
- Leong, J.S., Jantzen, S.G., von Schalburg, K.R., Cooper, G.A., Messmer, A.M., Liao, N.Y., Munro, S., Moore, R., Holt, R.A., Jones, S.J., Davidson, W.S., Koop, B.F., 2010. Salmo salar and Esox lucius full-length cDNA sequences reveal changes in evolutionary pressures on a post-tetraploidization genome. BMC Genomics 11, 279.
- pressures on a post-tetraploidization genome. BMC Genomics 11, 279. Lerner, D.T., Bjornsson, B.T., McCormick, S.D., 2007. Aqueous exposure to 4nonylphenol and 17 beta-estradiol increases stress sensitivity and disrupts ion regulatory ability of juvenile Atlantic salmon. Environmental Toxicology and Chemistry 26, 1433–1440.
- Liber, K., Gangl, J.A., Corry, T.D., Heinis, L.J., Stay, F.S., 1999. Lethality and bioaccumulation of 4-nonylphenol in bluegill sunfish in littoral enclosures. Environmental Toxicology and Chemistry 18, 394–400.
- Liu, H., Takano, T., Peatman, E., Abernathy, J., Wang, S.L., Sha, Z.X., Kucuktas, H., Xu, D.H., Klesius, P., Liu, Z.J., 2010. Molecular characterization and gene expression of the channel catfish ferritin H subunit after bacterial infection and iron treatment. Journal of Experimental Zoology Part A: Ecological Genetics and Physiology 313A, 359–368.
- Luckenbach, J.A., Iliev, D.B., Goetz, F.W., Swanson, P., 2008. Identification of differentially expressed ovarian genes during primary and early secondary oocyte growth in coho salmon, *Oncorhynchus kisutch*. Reproductive Biology and Endocrinology, 6.
- Madsen, S.S., Mathiesen, A.B., Korsgaard, B., 1997. Effects of 17beta-estradiol and 4-nonylphenol on smoltification and vitellogenesis in Atlantic salmon (Salmo salar). Fish Physiology and Biochemistry 17, 303–312.
- Madsen, S.S., Skovbolling, S., Nielsen, C., Korsgaard, B., 2004. 17-beta estradiol and 4-nonylphenol delay smolt development and downstream migration in Atlantic salmon, *Salmo salar*. Aquatic Toxicology 68, 109–120.
- Masuyama, H., Hiramatsu, Y., Kunitomi, M., Kudo, T., MacDonald, P.N., 2000. Endocrine disrupting chemicals, phthalic acid and nonylphenol, activate Pregnane X receptor-mediated transcription. Molecular Endocrinology 14, 421–428.
- McCormick, S.D., 1993. Methods for nonlethal gill biopsy and measurement of Na+ K+-ATPase activity. Canadian Journal of Fisheries and Aquatic Sciences 50, 656–658.
- McCormick, S.D., 2001. Endocrine control of osmoregulation in teleost fish. American Zoologist 41, 781–794.
- McCormick, S.D., Bjornsson, B.T., Sheridan, M., Eilertson, C., Carey, J.B., Odea, M., 1995. Increased daylength stimulates plasma growth-hormone and gill Na⁺, K⁺ ATPase in Atlantic Salmon (*Salmo salar*). Journal of Comparative Physiology B: Biochemical Systemic and Environmental Physiology 165, 245–254.
- McCormick, S.D., O'Dea, M.F., Moeckel, A.M., Lerner, D.T., Bjornsson, B.T., 2005. Endocrine disruption of parr-smolt transformation and seawater tolerance of Atlantic salmon by 4-nonylphenol and 17 beta-estradiol. General and Comparative Endocrinology 142, 280–288.
- McCormick, S.D., Shrimpton, J.M., Moriyama, S., Bjornsson, B.T., 2007. Differential hormonal responses of Atlantic salmon parr and smolt to increased daylength: a possible developmental basis for smolting. Aquaculture 273, 337–344.

- Munro, C., Stabenfeldt, G., 1984. Development of a microtitre plate enzyme immunoassay for the determination of progesterone. Journal of Endocrinology 101, 41–49.
- Naylor, C.G., Mieure, J.P., Adams, W.J., Weeks, J.A., Castaldi, F.J., Ogle, L.D., Romano, R.R., 1992. Alkylphenol ethoxylates in the environment. Journal of the American Oil Chemists Society 69, 695–703.
- Neves, J.V., Wilson, J.M., Rodrigues, P.N.S., 2009. Transferrin and ferritin response to bacterial infection: the role of the liver and brain in fish. Developmental and Comparative Immunology 33, 848–857.
- Nilsen, T.O., Ebbesson, L.O., Madsen, S.S., McCormick, S.D., Andersson, E., Bjornsson, B.T., Prunet, P., Stefansson, S.O., 2007. Differential expression of gill Na⁺, K⁺-ATPase alpha- and beta-subunits, Na⁺, K⁺, 2Cl⁻ cotransporter and CFTR anion channel in juvenile anadromous and landlocked Atlantic salmon Salmo salar. Journal of Experimental Biology 210, 2885–2896.
- Parrish, D.L., Behnke, R.J., Gephard, S.R., McCormick, S.D., Reeves, G.H., 1998. Why aren't there more Atlantic salmon (*Salmo salar*)? Canadian Journal of Fisheries and Aquatic Sciences 55, 281–287.
- Peatman, E., Baoprasertkul, P., Terhune, J., Xu, P., Nandi, S., Kucuktas, H., Li, P., Wang, S.L., Somridhivej, B., Dunham, R., Liu, Z.J., 2007. Expression analysis of the acute phase response in channel catfish (Ictalurus punctatus) after infection with a Gram-negative bacterium. Developmental and Comparative Immunology 31, 1183–1196.
- Rajeevan, M.S., Vernon, S.D., Taysavang, N., Unger, E.R., 2001. Validation of arraybased gene expression profiles by real-time (kinetic) RT-PCR. Journal of Molecular Diagnostics 3, 26–31.
- Robertson, L.S., McCormick, S.D. Transcriptional profiling of the parr-smolt transformation in Atlantic salmon, unpublished results.
- Ruggeri, B., Ubaldi, M., Lourdusamy, A., Soverchia, L., Ciccocioppo, R., Hardiman, G., Baker, M.E., Palermo, F., Polzonetti-Magni, A.M., 2008. Variation of the genetic expression pattern after exposure to estradiol-17 beta and 4-nonylphenol in male zebrafish (*Danio rerio*). General and Comparative Endocrinology 158, 138–144.
- Saeed, A.I., Sharov, V., White, J., Li, J., Liang, W., Bhagabati, N., Braisted, J., Klapa, M., Currier, T., Thiagarajan, M., Sturn, A., Snuffin, M., Rezantsev, A., Popov, D., Ryltsov, A., Kostukovich, E., Borisovsky, I., Liu, Z., Vinsavich, A., Trush, V., Quackenbush, J., 2003. TM4: a free, open-source system for microarray data management and analysis. Biotechniques 34, 374–378.
- Schroeder, A., Mueller, O., Stocker, S., Salowsky, R., Leiber, M., Gassmann, M., Lightfoot, S., Menzel, W., Granzow, M., Ragg, T., 2006. The RIN: an RNA integrity number for assigning integrity values to RNA measurements. BMC Molecular Biology 7, 3.
- Sohoni, P., Sumpter, J.P., 1998. Several environmental oestrogens are also antiandrogens. Journal of Endocrinology 158, 327–339.
- Storey, J.D., Tibshirani, R., 2003. Statistical significance for genomewide studies. Proceedings of the National Academy of Sciences of the United States of America 100, 9440–9445.
- von Schalburg, K.R., Rise, M.L., Cooper, G.A., Brown, G.D., Gibbs, A.R., Nelson, C.C., Davidson, W.S., Koop, B.F., 2005. Fish and chips: various methodologies demonstrate utility of a 16,006-gene salmonid microarray. BMC Genomics, 6.
- Wang, W., Zhang, M., Sun, L., 2011. Ferritin M of Cynoglossus semilaevis: an iron-binding protein and a broad-spectrum antimicrobial that depends on the integrity of the ferroxidase center and nucleation center for biological activity. Fish and Shellfish Immunology 31, 269–274.
- Watanabe, H., Suzuki, A., Goto, M., Lubahn, D.B., Handa, H., Iguchi, T., 2004. Tissuespecific estrogenic and non-estrogenic effects of a xenoestrogen, nonylphenol. Journal of Molecular Endocrinology 33, 243–252.
- White, R., Jobling, S., Hoare, S.A., Sumpter, J.P., Parker, M.G., 1994. Environmentally persistent alkylphenolic compounds are estrogenic. Endocrinology 135, 175–182.
- Yadetie, F., Male, R., 2002. Effects of 4-nonylphenol on gene expression of pituitary hormones in juvenile Atlantic salmon (*Salmo salar*). Aquatic Toxicology 58, 113–129.
- Zhang, X., Wei, W., Wu, H.Z., Xu, H.Y., Chang, K.M., Zhang, Y.X., 2010. Gene cloning and characterization of ferritin H and M subunits from large yellow croaker (*Pseudosciaena crocea*). Fish and Shellfish Immunology 28, 735–742.